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# Catalytic Decomposition of N<sub>2</sub>O: Best Achievable Methods and Processes

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Available online	and the use of memorane systems in the separation to molecules $N_2$ and $\Omega$ , were studied too. Decomposition of N $\Omega$ has been studied in
March 11, 2013	and $O_2$ were studied, too. Decomposition of $N_2O$ has been studied in
Keywords:	the temperature rate of 6/3 to 8/3 K over supported catalysts of
N <sub>2</sub> O decomposition;	chemical elements: Pd, Rh, Ru, Ni, Pt, Zn, Fe, Cu, Ir, over $\gamma$ - Al <sub>2</sub> O <sub>3</sub>
split N <sub>2</sub> O;	showing their best catalytic activity. M-zeolites, (M = Cu, Fe, Co,
catalytic zeolite systems;	etc.) supported on perovskite or precious metals such as Pd, Rh
cobalt spinels;	zeolites and dominant iron and copper catalytic structures such as
Rh catalytic systems;	ZSM-5, MFI, BEA, BETA investigated in the temperature rate of 583
Cu catalytic systems;	to 775 K with best catalytic activity. Iron zeolites are more prevalent
membrane $N_2O$ separation.	at high concentrations showing good catalytic behavior only at high
•	tomporatures. The spinal catalyst $T_{\rm Dec}$ ( $C_{\rm Dec}$ ) and $Ph/Mullite$
	1000000000000000000000000000000000000
	catalyst offer up to complete $N_2O$ conversion.
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## 1. Introduction

It is well known that nitrous oxide is considered a dangerous environmental pollutant because it contributes to the destruction of stratospheric ozone and simultaneously is also considered as a

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greenhouse gas. The concentration of  $N_2O$  in the atmosphere increases every year up to 0.2–0.3%, and this increase is caused mainly by anthropogenic activities. More specific, chemical processes related to the production and use of nitric acid and fluidized bed combustion are the two main nitrous oxide sources, and their contribution to the total nitrous oxide emissions amounts up to 20%. Minimization of N<sub>2</sub>O emission can be accomplished mainly by two ways, either by lowering the formation of N<sub>2</sub>O or after treatment where catalysis introduces the N<sub>2</sub>O abatement by the direct decomposition into nitrogen and oxygen. Metal such as transition metal oxides, zeolites and noble metals, especially rhodium, dispersed on oxide supports exhibit catalytic activity in N<sub>2</sub>O decomposition. [1]

Nitrous oxide ( $N_2O$ ), for many years has not received the attention of scientific interest. But, in recent years, due to the significant impact of greenhouse gases in the atmosphere and the increasing annual concentration a research activity has started, which involves various processes and technologies of reduction of it and splitting into harmless molecules,  $N_2$  and  $O_2$ .

A well understandable picture of the global nitrogen cycle is shown in Figure 1. (1) urban areas:  $NO_x$  is produced from high-temperature combustion. (2) Biomass burning:  $NO_x$ ,  $NH_3$ , and  $N_2O$  are produced from combustion of plant materials. (3) Agriculture: nitrogen is fixed through fertilizer manufacturing and the planting of nitrogen-fixing crops. Denitrification produces  $N_2$  and  $N_2O$ .  $NH_3$  is produced from fertilizer use and livestock excreta. (4) Natural fluxes: nitrogen fixing microbes in soil and sea water supply plants with usable nitrogen. Denitrification releases  $N_2$  and  $N_2O$  to atmosphere. (5) Wet and dry deposition: Nitrate and ammonia are removed from the atmosphere to both land and sea. (6) Lightning:  $NO_x$  is produced via high-temperature reaction of  $N_2$  and  $O_2$ . [2]

The following mechanisms is proposed for the direct N<sub>2</sub>O decomposition:

- (a)  $N_2O + (*) \rightarrow N_2 + *O$  $N_2O + *O \rightarrow N_2 + O*O$  $O*O \rightarrow O_2 + (*)$
- (b)  $2N_2O + 2(*) \rightarrow 2N_2 + 2*O$  $2*O \rightarrow O_2 + 2(*)$
- (c)  $N_2O + (*) \rightarrow N_2 + *O$

$$N_2O + *O \rightarrow N_2 + *O_2$$
$$N_2O + *O_2 \rightarrow N_2 + *O_3$$
$$*O_3 \rightarrow *O + O_2$$



Figure 1: Nitrogen cycle [2].

In mechanism (b) (Langmuir-Hinshelwood mechanism), the migration of oxygen from one active site (\*) followed by recombination with another oxidized site is the rate-determining step. This reaction mechanism requires the active participation of, at least, two iron centers that are not necessarily located in adjacent positions. In mechanisms (a) and (c) (Eley–Rideal mechanisms), N<sub>2</sub>O decomposition and oxygen evolution occur at the same isolated sites (\*) after successive collisions between N<sub>2</sub>O with (\*) and \*O for (a), and with (\*), \*O, and \*O<sub>2</sub> for (c). The proposed above mechanisms are all in agreement with transient response experiments, showing that N<sub>2</sub> appears before O<sub>2</sub> upon direct N<sub>2</sub>O decomposition pulses in the 773 to 848 K temperature range. This is happens because in all cases the global decomposition reaction is limited by the reactions steps leading to gas O<sub>2</sub>. More specific, recently accurate temporal analysis of products strongly implies that mechanism (a) is the most likely. On the other hand, from the point of view of the oxygen species formed on iron active sites (\*), the three mechanisms are associated with increasingly complex oxygen species: \*O, O\*O, and \*O<sub>2</sub> for (a); \*O for (b); and \*O, \*O<sub>2</sub> and Corresponding author (Z. Ziaka). Tel/Fax: +30-2310-275473 E-mail addresses: © 2013. American Transactions on bookeng@hotmail.com, z.ziaka@ihu.edu.gr.

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\*O<sub>3</sub> for (c). [3]

The anthropogenic sources of N<sub>2</sub>O emissions from the Annex I group are differentiated, as shown in Table 1, which is based on UNFCCC data (UNFCCC, 2009b). The energy and industrial processes of N<sub>2</sub>O emission sources contributed to ca.12% and 10% of total N<sub>2</sub>O emissions in the year 2007, respectively. If we compare the net emissions between the years 1990 and 2007, the N<sub>2</sub>O emissions generated by agriculture are decreased significantly over that period, but yet the extent of reduction was much higher in industrial processes than in the agricultural sector. As an explanation, one can say that it is possible to accomplish such a high degree of reduction in the volume of emissions generated by industrial processes because the N<sub>2</sub>O produced by those sources is likely destructed by using conventional N<sub>2</sub>O decomposition technologies. [4]

Sources	1990 Net emission <sup>a</sup>	2007 Net emission <sup>a</sup>	Difference of emission <sup>b</sup>	Reduction <sup>c</sup> %
Energy	119,665	113,535	6130	5
Industrial processes	184,387	91,983	92,404	50
Solvent and other product use	10,994	9,863	1081	10
agriculture	895,550	695,811	199,739	22
LULUCI <sup>d</sup>	10,825	14,367	-3542	-33
Waste	28,005	31,120	-3115	-11
other	32	33	-1	-3
Total	1,249,408	956,712	292,696	23

 Table 1: Net emissions of N2O from the anthropogenic sources by members of the Annex I group (UNFCCC, 2009b)

<sup>a</sup> Annual emission in terms of 1000 tons CO<sub>2</sub> equivalent.

<sup>b</sup> Net emission in 1990- net emission in 2007.

<sup>c</sup> (Net emission in 1990- net emission in 2007)/ net emission in 1990 x 100.

<sup>d</sup>Land use, land-use change and forestry.

Moreover, the number of clean development mechanism (CDM) projects aimed at reducing N<sub>2</sub>O emissions has increased the last years. Although N<sub>2</sub>O reduction projects account for only 2.6% of all CDM projects, these N<sub>2</sub>O reduction projects account for 13% of the total reduction of all greenhouse gases measured on a CO<sub>2</sub> equivalent basis regarding the CDM. Specifically, China is the host nation for half of all N<sub>2</sub>O CDM reduction projects while approximately 78% of the N<sub>2</sub>O reduction technologies for these CDM projects come from Japan, the UK, and Switzerland. N<sub>2</sub>O reduction in CDM projects has been applied to production plants for such chemicals as nitric acid, adipic acid, and caprolactam. The relevant technologies for N<sub>2</sub>O reduction with the selection of an

appropriate reduction strategy dependent on the specifications of the production process.

Thirty-nine case studies for the N<sub>2</sub>O decomposition have been investigated and examined [45], and in this paper among them we present the best catalytic application of the N<sub>2</sub>O decomposition through the catalytic systems below including [37], [38] [39], [40]:

- Split N<sub>2</sub>O on Supported Catalysts
- Split N<sub>2</sub>O on Zeolite catalysts
- Split N<sub>2</sub>O over cobalt Spinels
- Split N<sub>2</sub>O over rhodium catalyst
- Split N<sub>2</sub>O over Cu Catalysts

In addition, it has been also studied the case of N<sub>2</sub>O separation into nitrogen and oxygen via photocatalysis ( [41], [42] ) and via membranes (six case studies), [34].

Moreover, nitrogen gas,  $N_2$ , is the most abundant gas in the atmosphere at 78.1% by volume. In terms of physical properties, Nitrogen is a colorless gas at room temperature. Although  $N_2$  is generally quite inert chemically, it is utilized by nitrogen-fixing organisms and this is the primary natural nitrogen input to terrestrial and marine ecosystems. However, biological nitrogen fixation refers to the natural process that is performed by a variety of bacteria and algae, both symbiotic and free living. On the other side, industrial nitrogen fixation refers to the industrial production of NH<sub>3</sub> and nitrates from  $N_2$ , mainly for fertilizers. In addition, nitrogen-fixing organisms are resided on the roots of many leguminous plants (clover, soybeans, chickpeas, etc.) and have been utilized agriculturally as a mean of replenishing soil nitrogen ("green manures").

The natural flux of  $N_2$  from the atmosphere to oceans and terrestrial ecosystems is balanced by an almost equal source of  $N_2$  to the atmosphere in a process called denitrification. In denitrification, certain microbes utilize aqueous  $NO_3^-$  ion as an oxidant in low-oxygen environments. Thus, the combined effects of natural biological nitrogen fixation and denitrification result in a nearly balanced cycle with respect to the atmospheric  $N_2$ . Denitrification releases a smaller amount of  $N_2O$ , as well.

As mentioned earlier, Nitrous oxide, or N<sub>2</sub>O, is a colorless, fairly unreactive gas at room

temperature. It is also the second most abundant nitrogen compound in the atmosphere with a current (year 2000) average global mixing ratio of 316 parts per billion by volume (ppbv). Because of its low reactivity and low water solubility,  $N_2O$  has a lifetime of about 100 years in the atmosphere and is well mixed throughout the global troposphere. Due the fact that it has no significant sinks in the troposphere,  $N_2O$  eventually gets transported to the stratosphere where it can react or photolyze. The products of this decomposition demonstrate an important parameter in the chemistry of the stratosphere. In denitrification, 80 to 100% of the nitrogen released is in the form of  $N_2$ , with the remainder being  $N_2O$ , although under certain environmental conditions, nitrous oxide can end up as a major end product.

Although denitrification is a natural process, there is an important increase in denitrification and  $N_2O$  release coming from the application of nitrogenous fertilizers, which is indeed a certain contributor to the global increase in observed  $N_2O$  concentrations. Moreover,  $N_2O$  is produced as well from biomass burning.

Because of these additional anthropogenic contributions,  $N_2O$  sources are now higher than the sinks and  $N_2O$  is accumulating in the atmosphere. Because  $N_2O$  is a greenhouse gas and shows an important impact in the stratospheric ozone depletion, considering the contributions from agricultural releases and biomass burning of  $N_2O$  and, if possible, minimizing these fluxes are new areas of further investigation.

Table 2 shows the global average lifetime, global average mixing ratio or the range of observed mixing ratios, and an estimate of the global burden for some of the key nitrogen species. Long-lived compounds, such as  $N_2$  and  $N_2O$ , have very uniform distributions throughout the globe. This helps us to determine easily their total abundance and average global concentration based on a relatively small number of observations. However, short-lived species, such as  $N_x$  and  $NH_3$ , have highly inhomogeneous distributions, which make it much more difficult to determine their global abundance. For example, the  $NO_x$  mixing ratios vary over 5 orders of magnitude between urban and remote regions.

Furthermore, for both  $NO_x$  and  $NH_3$ , the lifetime is dependent on local conditions. For  $NO_x$ , the lifetime in urban air due to reaction with OH is about half a day, whereas in the upper troposphere it can be many days, depending on the OH molecular density. So for both  $NO_x$  and

 $NH_3$ , the global burdens have high uncertainties, on the order of  $\pm 100\%$ . For  $N_2O$ , there is good evidence that modern-day mixing ratios are larger than pre-industrial values by 12–15%. This outcome is based on direct atmospheric observations over the past two decades and ice core data extending back several thousand years.

Figure 2 shows recent atmospheric data, collected at four background stations by the National Oceanographic and Atmospheric Administration's Climate Monitoring and Diagnostics Laboratory (NOAA–CMDL). Besides, it is shown an estimate of the pre-industrial mixing ratio of  $N_2O$ , as determined by ice core data. For  $NH_3$  and  $NO_x$ , the global burden has certainly increased significantly as a result of human contribution; however, due to the highly variable nature of these reactive gases, it is difficult to quantify the exact amounts. [2]

 Table 2: Mixing ratios and burdens of atmospheric nitrogen compounds.
 The N<sub>2</sub>O mixing ratio is rising by 0.7 ppbv per year. These values are for the year 2000.

compound	lifetime	global mean value	global burden (tgN)
N <sub>2</sub>	Millions of	78.08%	3.9 x 10 <sup>9</sup>
	years		
N <sub>2</sub> O	~100 years	316 ppbv	1600
NH <sub>3</sub>	~3 years	0.05-50 ppbv	~0.4
NO <sub>X</sub>	$\sim 1 \text{ day}$	0.005-200 ppbv	~0.2



**Figure2:** Global average N<sub>2</sub>O mixing ratio (ppbv) from recent atmospheric data and estimated from ice core data. The NOAA-CMDL data are based on observations made at four background stations [2].

After the Kyoto Protocol regulation for the emission of six green-house gases (CO2, CH4,

N<sub>2</sub>O, HFCs, PFCs and SF<sub>6</sub>), a number of clean development mechanism (CDM) projects has been registered under the United Nations Framework Convention on Climate Change (UNFCCC). CDM projects have been developed to assist the reduction of greenhouse gas emissions in developing countries included in the non-Annex I group, by allowing them to earn certified emission reduction (CER) credits which can be traded and used by industrialized countries in the Annex I group. Since the CDM targets the reduction of greenhouse gas (GHG) emissions in developing countries, an acceleration of technology transfer has been introduced from developed countries to the developing ones. Moreover, CDM projects stimulate the development of reduction technology to obtain more CER credits. Furthermore, emissions by greenhouse gas sources have been reported in order to systemize them for CDM projects.

Table 5. 1 (et ennission,	s of each greenhou	ise gases by member	s of the Annex I grou	p(0111  cec, 20090)
Gas	Net emission <sup>a</sup>	Net emission <sup>a</sup>	Difference of	Reduction <sup>°</sup> %
Uas	1990	2007	emission <sup>b</sup>	
CO <sub>2</sub>	13,514,353	13,375,386	138,967	1
CH <sub>4</sub>	2,248,913	1,906,667	342,246	15.2
N <sub>2</sub> O	1,249,408	956,712	292,696	23.4
Aggregated F-gases <sup>d</sup>	267,111	308,379	-41,267	-5.4
Total	17,279,785	16,547,143	732,642	42

Table 3: Net emissions of each greenhouse gases by members of the Annex I group (UNFCCC, 2009b)

<sup>a</sup> Annual emission in terms of 1000 tons CO<sub>2</sub> equivalent.

<sup>b</sup>Net emission in 1990- net emission in 2007.

<sup>c</sup> (Net emission in 1990- net emission in 2007)/ net emission in 1990 x 100.

<sup>d</sup> HFCs, PFCs, SF<sub>6</sub>

Tables 1 and 3, show the net emission of each greenhouse gas by members of the Annex I group from the years 1990 to 2007 for comparison, based on data provided by the UNFCCC (UNFCCC, 2009a). Although CO<sub>2</sub> accounts for ca.81% of all net greenhouse gas emissions in 2007, the reduction of CO<sub>2</sub> is low in comparison with that of CH<sub>4</sub> and N<sub>2</sub>O. On the other hand, it has been presented that the reduction of N<sub>2</sub>O is quite high compared with other greenhouse gases, even though the amount by which N<sub>2</sub>O emissions has been reduced is at a lower degree than that of the CH<sub>4</sub>. For this reason, the reduction of N<sub>2</sub>O rather than other greenhouse gases is bound to attract more interest in CDM projects. The anthropogenic sources of N<sub>2</sub>O emissions from the Annex I group are varied, as shown in Tables 1 and 3, which is based on UNFCCC data (UNFCCC, 2009b). The energy and industrial processes of N<sub>2</sub>O emission sources contributed to ca.12% and 10% of total N<sub>2</sub>O emissions in the year 2007, respectively. If we compare net emissions between the years 1990 and 2007, N<sub>2</sub>O emissions generated by agriculture decreased considerably over that period, yet the extent of reduction is much higher in industrial processes than in the agricultural sector. It

was possible to accomplish such a high degree of reduction in the volume of emissions generated by industrial processes because, as it was addressed before, the N<sub>2</sub>O produced by those processes is likely destructed by using conventional N<sub>2</sub>O decomposition technologies. Thus, the Annex I group has tried to transfer the technologies to the non-Annex I group via the implementation of CDM projects. This paper investigates the present status and trends of CDM projects for N<sub>2</sub>O reduction, and proposes a possible future direction for CDM projects.

#### Present status of CDM projects to reduce N<sub>2</sub>O emissions

By December 2009, a total of 1952 CDM projects had been registered with the UNFCCC, from which 2.6% were related to N<sub>2</sub>O reduction, showing a sharp increase in the number of CDM projects aimed at N<sub>2</sub>O reduction since 2007 (see Figure 3) (UNFCCC, 2009c). Despite the small number of CDM projects contributing to the reduction of the N<sub>2</sub>O emissions, it is worthy to be addressed that the reduction of N<sub>2</sub>O emissions in terms of their CO<sub>2</sub> equivalence accounts for 13% of the total reduction of all greenhouse gases. There are 12 countries that host CDM projects for N<sub>2</sub>O reduction, namely China, Uzbekistan, Republic of Korea, Israel, South Africa, Brazil, India, Chile, Colombia, Egypt, Pakistan and Thailand. China and Uzbekistan have registered 50% and 11.5% of the total number of the CDM projects, respectively (see Figure 4a) (UNFCCC, 2009c). For the CDM projects involving the reduction of N<sub>2</sub>O, eight nations –the U.K., Japan, Switzerland, Germany, France, the Netherlands, Canada and Austria have considered the host parties with new technologies for N<sub>2</sub>O reduction. The U.K., Japan and Switzerland have contributed 32.4%, 26.7% and 19.7% of the technology transfer, respectively (see Figure 4b) (UNFCCC, 2009c). The transferred technologies were employed to reduce N<sub>2</sub>O emitted by industrial processes involved in the production of chemicals such as nitric acid, adipic acid and caprolactam. With respect to the CDM projects, the applied industrial process consists of 90% nitric acid, 7.8% adipic acid and 1.9% caprolactam, as shown in Figure 5 (UNFCCC, 2009c). The large number of projects target in nitric acid production may be contributed to the application of various technologies to N<sub>2</sub>O reduction. The technical measures for baseline and monitoring in the CDM projects are described in methodologies of the CDM projects. [4]



Figure 3: The annual number of CDM projects for N<sub>2</sub>O reduction (UNFCCC, 2009c) [4]



Figure 4: The number of CDM projects for  $N_2O$  reduction a) the host parties, b) the technology suppliers (UNFCCC, 2009c) [4].



Figure 5: The number of CDM projects for  $N_2O$  reduction in industrial processes (UNFCCC, 2009c) [4].

#### The proposed mechanism in direct N<sub>2</sub>O decomposition

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The most popular mechanisms in direct N<sub>2</sub>O decomposition are presented as follows:

(a) 
$$N_2O + (*) \rightarrow N_2 + *O$$
  
 $N_2O + *O \rightarrow N_2 + O*O$   
 $O*O \rightarrow O_2 + (*)$   
(b)  $2N_2O + 2(*) \rightarrow 2N_2 + 2*O$   
 $2*O \rightarrow O_2 + 2(*)$   
(c)  $N_2O + (*) \rightarrow N_2 + *O$   
 $N_2O + *O \rightarrow N_2 + *O_2$   
 $N_2O + *O_2 \rightarrow N_2 + *O_3$   
 $*O_3 \rightarrow *O + O_2$ 

In mechanism (b) (Langmuir–Hinshelwood mechanism), the migration of oxygen from one active site (\*) followed by recombination with another oxidized site is the rate-determining step. This proposed reaction mechanism implies the active participation of at least two iron centers that are not necessarily located in adjacent positions. In mechanisms (a) and (c) (Eley–Rideal mechanisms), N<sub>2</sub>O decomposition and oxygen evolution happen at the same isolated sites (\*) after successive collisions between N<sub>2</sub>O with (\*) and \*O for (a), and with (\*), \*O, and \*O<sub>2</sub> for (c). The above analyzed mechanisms are all in a good agreement with transient response experiments, indicating that N<sub>2</sub> appears before O<sub>2</sub> upon direct N<sub>2</sub>O decomposition pulses in the 773 to 848 K temperature range. This happens because in all cases the global decomposition reaction is limited by the reaction steps leading to O<sub>2</sub> gas phase.

More recent accurate temporal analysis of products strongly implies that mechanism (a) is the most likely. On the other hand, from the point of view of the oxygen species formed on iron active sites (\*), the three mechanisms are associated with increasingly complex oxygen species: \*O and O\*O for (a); \*O for (b); and \*O,  $*O_2$ , and  $*O_3$  for (c).

#### 2. Study Details and Discussion

#### **Best case study 1:** N<sub>2</sub>O Abatement Over *y*-Al<sub>2</sub>O<sub>3</sub> Supported Catalysts.

Numerous catalytic systems, such as supported or unsupported noble metals, pure oxides,

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mixed oxides, are reported to be active for N<sub>2</sub>O decomposition/ reduction reactions [5], [6], [7].

#### Catalyst.

Through their study in 2009, G. Pekridis et al. [7], investigated the effect of excess oxygen and reducing agents (CH<sub>4</sub> and C<sub>3</sub>H<sub>8</sub>) on N<sub>2</sub>O decomposition, over a series of noble and transition metal catalysts (Pd, Rh, Ru, Cu, Fe, In and Ni), supported on  $\gamma$ -Al<sub>2</sub>O<sub>3</sub> carrier. It was noticed that N<sub>2</sub>O decomposition rate was inhibited by oxygen, whereas the catalytic activity was generally increased in the presence of reducing agents, with C<sub>3</sub>H<sub>8</sub> to be more efficient than CH<sub>4</sub>. [7]

#### Materials and Reaction System.

A series of Pd-, Rh-, Ru-, Cu-, Fe-, In-, and Ni-based on alumina catalysts were prepared using the dry impregnation method. Water solutions of the corresponding metal salts were impregnated on  $\gamma$ -Al<sub>2</sub>O<sub>3</sub> carrier (Engelhard, 180–355 µm) so as to yield 2.0 wt% metal loading.

After impregnation took place, the catalysts were dried at 120 °C for 2 h and then were calcined at 600 °C for 4 h under air flow. The actual metal content was measured using ICP-AES, while the BET surface areas of the samples were measured at 77 K with an Autosorb-1 apparatus.

All the catalysts prepared in the present work are listed in Table 4 along with the corresponding precursor materials used, their actual metal loading, and BET surface areas.

Type of Catalyst	Precursor		Metal loading (%w/w)	Surface area (m <sup>2</sup> /g)
		Nominal	ICP measured	
γ-Al <sub>2</sub> O <sub>3</sub>	-	-	-	209
$Pd/\gamma$ - $Al_2O_3$	$Pd(NO_3)_2.2H_2O$	2	$2.078 \pm 0.075$	173
Rh/y-Al <sub>2</sub> O <sub>3</sub>	RhCl <sub>3</sub> .3H <sub>2</sub> O	2	$1.917 \pm 0.045$	170
$Ru/\gamma$ - $Al_2O_3$	RuCl.H <sub>2</sub> O	2	Not measured	192
Cu/γ-Al <sub>2</sub> O <sub>3</sub>	$Cu(NO_3)_2.3H_2O$	2	$1.992 \pm 0.060$	175
Fe/y-Al <sub>2</sub> O <sub>3</sub>	$Fe(NO_3)_3.9H_2O$	2	$1.960 \pm 0.050$	171
$In/\gamma - Al_2O_3$	$In(NO_3)_3.5H_2O$	2	$2.090 \pm 0.070$	184
$Ni/\gamma$ -Al <sub>2</sub> O <sub>3</sub>	$Ni(NO_3)_2.6H_2O$	2	$1.925 \pm 0.050$	129

**Table 4:** List of prepared catalysts.



Figure 6: Effect of Catalyst Entity and Gas Phase Oxygen on the N<sub>2</sub>O Decomposition [7].

#### Analysis.

Analyzing the experimental observations, the current results show that  $N_2O$  decomposition is the major pathway for  $N_2O$  abatement even in the presence of hydrocarbons; it works mainly as oxygen scavengers reducing the metal active sites. The different observed efficiencies of  $C_3H_8$  and  $CH_4$  can be explained by taking into account their different interaction with the catalyst's surface and consequently their capability to take away surface oxygen ad-atoms originated either from  $N_2O$  decomposition and/or gas phase oxygen dissociative adsorption.

Pure  $\gamma$ -alumina and alumina-supported catalysts were initially tested for the N<sub>2</sub>O decomposition to N<sub>2</sub> (the only N-containing product) in the absence of oxygen at the temperature range of 200 – 600°C.

Following, all catalysts were inhibited by the presence of oxygen in the gas phase, as oxygen is dissociatively adsorbed on the catalyst's surface to yield strongly chemisorbed oxygen species, identical to those produced by the  $N_2O$  decomposition. As a result a monoatomic oxygen layer is formed, hindering the reaction. The effect of catalyst entity and gas phase oxygen on the  $N_2O$  decomposition is shown in the Figure 6:

The effect of reducing agent on the  $N_2O$  reduction at lean conditions is shown in a combined set in Figure 7 [7].



Figure 7: The effect of reducing agent on the N<sub>2</sub>O reduction at Lean Conditions [7].



Figure 8: Effect of CO, H<sub>2</sub>O and SO<sub>2</sub> on the Activity and Stability of Pd/Al<sub>2</sub>O<sub>3</sub>[7].

The presence of reducing agent is going to accelerate the removal of oxygen ad-atoms from the catalyst surface, originated either by N<sub>2</sub>O decomposition or gas phase oxygen dissociative adsorption; this in turn can lead to enhanced catalytic activity at lower operation temperatures. In order to be comparable, the feed concentration of reducing agents was appropriately selected in order to obtain a constant C/N<sub>2</sub>O ratio. Therefore, propane concentration was kept constant to 0.17% (Figure 7), while the equivalent concentration for CH<sub>4</sub> equaled to 0.5% (Figure 7). It is clearly shown that the presence of reducing agents enhanced the catalytic activity under lean reaction conditions, with  $C_3H_8$  being more efficient than CH<sub>4</sub>, especially at lower temperatures. Another major agent that reduces the catalytic activity is the presence of automotive exhausts

containing several other gases leading in deactivation of the catalyst. Figure 8 is presenting a picture at 350 °C, of the achieved N<sub>2</sub>O conversion over Pd/Al<sub>2</sub>O<sub>3</sub> (one of the most efficient catalyst in this study) for various feed compositions that contained CH<sub>4</sub>, C<sub>3</sub>H<sub>8</sub>, CO, SO<sub>2</sub> and H<sub>2</sub>O. It is clear that the addition of a reducing agent increases the activity. While in the cases of CH<sub>4</sub> and C<sub>3</sub>H<sub>8</sub> the performance is slightly increased, CO seems to boost up the N<sub>2</sub>O conversion at much higher values. When SO<sub>2</sub> was added, the reaction is inhibited, especially in the case when C<sub>3</sub>H<sub>8</sub> was used as a reducing agent. This negative effect was also clear in the case of CO, however when CH<sub>4</sub> was employed this specific inhibition was limited.

Carbon monoxide definitely increases  $N_2O$  conversion, while  $H_2O$  and  $SO_2$  are clearly deactivated the catalyst. Thermal stability tests in the presence of  $SO_2$  over Pd–Al<sub>2</sub>O<sub>3</sub>, indicate a gradual irreversible decrease in activity, which was due to the formation of surface sulphates (Figure 9).



Figure 9: Effect of CO, H<sub>2</sub>O and SO<sub>2</sub> on the Activity and Stability of Pd/Al<sub>2</sub>O<sub>3</sub> [7].

#### Application.

Among all catalysts tested, noble metals (Pd, Ru, Rh) followed by  $Cu-Al_2O_3$  exhibit the best catalytic performance, independently of the experimental conditions imposed. In Figure 10 it is shown that a ca.100% conversion of N<sub>2</sub>O is achieved over Pd/Al<sub>2</sub>O<sub>3</sub> at 400 °C in the presence of O<sub>2</sub>. This is the best catalytic behavior over a series of noble and transition metal catalysts (Pd, Rh,

Ru, Cu, Fe, In and Ni), supported on  $\gamma$ -Al<sub>2</sub>O<sub>3</sub> carrier.



Figure 10: Best catalytic behavior [7].

Although Pd–Al<sub>2</sub>O<sub>3</sub> exhibited the best performance, its activity at low temperatures and  $H_2O + SO_2$  resistivity should be improved additionally (e.g., promotion with alkalis), in order to be considered as a potential solution for mobile exhaust gas treatment applications. [7]

#### Best case study 2: N<sub>2</sub>O decomposition over FeZSM-5.

In the last decade, attention was drawn to development of new methods for the abatement of  $N_2O$ . Transition metal-exchanged silicon-rich zeolites are excellent candidates for the decomposition of nitrous oxide to nitrogen and oxygen [8]-[15].

#### Catalyst.

The decomposition of  $N_2O$  on zeolite catalysts has been studied extensively. The large surface area and the simultaneous presence of Bronsted positions of the metal centers, contribute to the high performance of M-zeolites, where M = Cu, Fe, Co, etc., perovskite oxides and supported precious metals such as Pd, Rh.

Dominant zeolite systems are iron and copper on zeolite structures (ZSM-5, MFI, BEA, BETA).

In this case study in order to investigate the  $N_2O$  decomposition was used the Fe/ ZSM-5 catalyst.

#### Material and Reaction System.

H. Xia et al. investigated in 2009 the N<sub>2</sub>O decomposition through the Fe/ ZSM-5 catalyst, promoted by NO [10]. The experimental procedures were conducted on laboratory space, the gas is a mixture of pure  $N_2O$  and with impurities consisting of oxygen. The Fe/ZSM-5 catalysts were synthesized by solid-state ion-exchange method. The as-synthesized sample was calcined in O<sub>2</sub> flow at 823 K followed by the treatment in He flow at 1173 K for 2 h. The content of Fe in the catalyst is about 2.2 wt.%, corresponding to Fe/Al = 0.61. A single-pass plug-flow quartz reactor with an inner diameter of 4 mm was used to test the catalyst performance of Fe/ZSM-5 samples for  $N_2O$  decomposition. The catalyst sample (50 mg) was retained between two quartz wool plugs. Prior to N<sub>2</sub>O decomposition at a given temperature, the catalyst was pretreated in He flow at 1173 K for 1 h. During transient and steady-state catalytic measurements, the concentration of N<sub>2</sub>O was 5.0 vol.% and the gas hourly space velocity was kept at 24,000 ml/g h. For N<sub>2</sub>O decomposition in the presence of NO, the gas hourly space velocity was kept at 24,000 ml/g h and the concentrations of N<sub>2</sub>O and NO were 5 vol.% and 1 vol.%, respectively. It should be noted that the feeds contain trace amount of water (about 1-10 ppm). To identify the characteristics and the catalytic properties, the following methods were used: XRD, SEM, ICP, TPD, XANES, FTIR, EDX, BET, DRS, Mossbauer spectroscopy, UV-vis spectroscopy and Infrared spectroscopy.

For the best performance of the catalysts, nitric oxide boosters, silver, platinum, cobalt and the reducing agent ammonia (Selective Catalytic Reduction) were also used.

#### Analysis.

Figure 11 shows the decomposition of  $N_2O$  on Fe/ZSM-5 as a function of temperature in the absence and in the presence of NO.



**Figure 11:** Catalytic Decomposition of N2O in presence ( $\mathbf{V}$ ) and absence ( $\mathbf{I}$ ) of NO [10].

In a N<sub>2</sub>O/He feed, the catalyst presents a substantial N<sub>2</sub>O conversion only above 643 K. The presence of NO (NO/N<sub>2</sub>O = 0.2) significantly increases the activity of N<sub>2</sub>O decomposition. The N<sub>2</sub>O conversion curve is shifted to a temperature about 60 K lower for the same conversion. In addition, the addition of NO in the feed leads to a significant decrease in the apparent activation energy.

#### The role of NO as a promoter of the catalytic activity:

- The presence of NO regenerates the active sites of the catalyst and accelerates the reformation-(recombination) of oxygen during the catalytic decomposition of  $N_2O$ .
- plays an important role in the transport and storage of surface oxygen atoms and the fast desorption of  $O_2$ .
- helps the  $Fe^{+3}$  transformation to  $Fe^{+2}$ , releasing O<sub>2</sub> and H<sub>2</sub>O, which are the active sites of the catalyst on which the split of N<sub>2</sub>O takes place.

#### $N_2O$ decomposition on Fe/ZSM-5 in the absence and in the presence of NO

The experimental performance after  $N_2O$  decomposition reaches a steady-state. The desorption of  $O_2$  always delays more than that of  $N_2$ , but at lower temperature during  $N_2O$  decomposition; the desorption of  $N_2$  and  $O_2$  are almost synchronic in the presence of NO. This result unambiguously shows that the presence of NO enhances the desorption of  $O_2$  during  $N_2O$  direct decomposition.

The concentrations of  $N_2$  and  $O_2$  first reach to their maximum and then decay to a steady-state. Some surface oxygen atoms are deposited on the catalyst due to trace amount of water in the feed. After N<sub>2</sub>O decomposition reached a steady-state, a switch from N<sub>2</sub>O to NO is performed. In addition of methane and H<sub>2</sub>O, the addition of NO could remove hydroxyl groups bound to Fe species. The presence of NO removes the hydroxyl groups bound to the Fe species during the N<sub>2</sub>O direct decomposition.

As an indicative explanation for the role of NO in  $N_2O$  decomposition, schematic description of catalytic cycles is shown in the Figure 12.



Figure 12: Schematic description of the NO role in N<sub>2</sub>O decomposition [10].

The Figure 12 shows the results from in situ spectroscopy and transient response method study, for the first time. Bi-nuclear Fe sites are proposed to be the active sites for the NO-assisted  $N_2O$  decomposition.

In addition, the role of water in the structural transformation between the dehydroxylated and the hydroxylated bi-nuclear Fe sites was also elucidated. We conclude that the  $NO_2$  species adsorbed on binuclear Fe sites could be an intermediate species in  $N_2O$  decomposition in the presence of NO. The two adjacent  $NO_2$  adsorbed on the bi-nuclear Fe sites could react with each

other to release O<sub>2</sub>, or the adsorbed NO<sub>2</sub> species could react with N<sub>2</sub>O to release O<sub>2</sub>.

It has been presented that the structure of the Fe species in Fe/ZSM-5 is strongly dependent on the synthesis method, zeolite precursor, amount of Fe loading, and pretreatment conditions.

#### **Application**

The best performance (about 90%) in the decomposition of  $N_2O$  was observed by using the catalyst Fe-ZSM-5, at 583K (310°C) in the presence of NO.

Zeolites iron is more prevalent; however the problem is that at low concentrations of Fe they do not show good catalytic behavior, while, when the concentration is high, they show good catalytic activity only at high temperatures. The copper zeolites are promising catalytic systems in the reaction of  $N_2O$  decomposition and especially amorphous oxide systems that contain copper (based on CuO<sub>x</sub>). These are characterized by high mechanical and hydrothermal stability.

# Best case study 3: Excellent catalytic performance of $Zn_xCo_2O_4$ spinel catalysts for the decomposition of nitrous oxide.

Currently, oxide spinels of 3d transition metals have been shown an increasing fundamental and applied research activity because of their special electric, magnetic, and catalytic properties. Spinel oxides belong to a class of complex oxides with the chemical formulas of  $AB_2O_4$  in which A ions are generally divalent cations occupying tetrahedral sites and B ions are trivalent cations in octahedral sites. The cobalt oxide spinel  $Co_3O_4$ , for example, is the one receiving considerable interest [16]-[24]. Regarding their applications in catalysis, for example, it has been proved that the catalytic activity of these spinels depends essentially on two actors: the degree of substitution and the inversion degree of the spinel.

#### <u>Catalyst</u>

During the catalytic decomposition of N<sub>2</sub>O, using the spinel  $Zn_xCo_{1-x}Co_2O_4$ , the cobalt ions replace  $Co^{+2}$  ions with  $Zn^{+2}$ . The activity of the catalysts was found to depend on the degree of this substitution. L. Yan et al., investigated the role of this substitution and presented the best catalytic performance for the N<sub>2</sub>O decomposition until nowadays. [17]

#### **Materials and Reaction System**

The catalytic preparation was based on co-precipitation method, adding aqueous solution of  $K_2CO_3$  (15 wt%) in a mixture of aqueous solutions  $Co(NO_3)_2$  and Zn (NO<sub>3</sub>)<sub>2</sub> to produce the  $Zn_{0.36}Co_{0.64}Co_2O_4$  catalyst.

Before the catalytic experiments, 0.6 g catalyst was pretreated for 30 min by 20% oxygen in He at 400 °C in order to remove the organic compounds adsorbed on the surface of samples, then decreased to reaction temperature. The catalytic reaction was carried out in a standard fixed-bed flow reactor by passing a gaseous mixture of N<sub>2</sub>O (1000 ppm), 0 or 10 vol% O<sub>2</sub>, 0 or 5 vol% H<sub>2</sub>O in a He flow rate 100–150 ml/min to get a space velocity of 15,000 h<sup>-1</sup>. To get reliable N<sub>2</sub>O conversion, the reaction system was kept for 30 min at each reaction temperature, then it started to analyze the off-gases with two on-line gas chromatographs.

Table 5 shows the characterization results, the conversion and reaction rate of  $N_2O$  at 200  $^{\circ}C$  of the samples of different degree of  $Co^{2+}$  substitution by  $Zn^{2+}$ .

Catalyst formula <sup>a</sup>	BET surface area <sup>b</sup> $(m^2/g)$	Conversion of $N_2O^{c}$	Rate of $N_2O$ conversion <sup>b</sup> (x10 <sup>6</sup> mol/g s)				
Co <sub>3</sub> Co <sub>4</sub>	56.6	1.6	0.12				
Zn 0.36 Co 0.54Co2 O4	73.6	98.4	7.31				
Zn 0.54 Co 0.46Co2 O4	76.2	90.2	6.71				
Zn <sub>0.70</sub> Co <sub>0.30</sub> Co <sub>2</sub> O <sub>4</sub>	75.1	84.0	6.25				
Zn 0.98 Co 0.02 Co2 O4	72.8	60.1	4.46				
1.1% Au/ Co <sub>3</sub> O <sub>4</sub>	23.6	22.6	4.04				

 Table 5: The characterization data of the catalysts

<sup>a</sup> Determined by AAS method.

<sup>b</sup> Characterized with ASAP 2010.

<sup>c</sup> Determined at 200 °C.

#### <u>Analysis</u>

#### Catalytic decomposition activity for $N_2O$

The conversion of N<sub>2</sub>O to N<sub>2</sub> and O<sub>2</sub> over  $Zn_xCo_{1-x}Co_2O_4$  spinel oxides at 200°C (shown in Table 5) indicated that the introduction of Zn into the spinel structure of Co<sub>3</sub>O<sub>4</sub> led to a significant improvement (conversion of N<sub>2</sub>O from 1.6% over pure Co<sub>3</sub>O<sub>4</sub> to 98.4% over  $Zn_{0.36}Co_{0.64}Co_2O_4$  at

200°C) in catalytic activity for N<sub>2</sub>O decomposition. Then with increasing the x value, the catalytic activity of samples decreased slowly (conversion of N<sub>2</sub>O decreased from 98.4% at x = 0.36 to 60.5% at x = 0.98).

On the weight basis of catalyst, the  $Zn_xCo_{1-x}Co_2O_4$  catalysts gave much higher activity than that of pure Co<sub>3</sub>O<sub>4</sub>. Moreover, on the basis of surface area of catalysts,  $Zn_xCo_{1-x}Co_2O_4$  samples still are more active. For instance, when comparing the N<sub>2</sub>O conversion of the  $Zn_{0.36}Co_{0.64}Co_2O_4$ sample with conversion of pure Co<sub>3</sub>O<sub>4</sub> at 200 °C, it is clear that the specific activity (per square meter of catalyst) is still much higher over the  $Zn_{0.36}Co_{0.64}Co_2O_4$  catalyst. Compared to the rate (Table 5) of N<sub>2</sub>O decomposition of 1.1% Au/Co<sub>3</sub>O<sub>4</sub>, the rate of  $Zn_{0.36}Co_{0.64}Co_2O_4$  is higher, which indicated that the  $Zn_xCo_{1-x}Co_2O_4$  catalysts were more active than that of 1.1% Au/Co<sub>3</sub>O<sub>4</sub> for direct N<sub>2</sub>O decomposition reaction.



**Figure 13:** The effect of gas composition on N<sub>2</sub>O conversion at the space velocity of 15,000 h<sup>-1</sup>. Gas composition ( $\Diamond, \blacklozenge$ ), 1000 ppm N<sub>2</sub>O; ( $\Box, \blacksquare$ ), 1000 ppm N<sub>2</sub>O+10% O<sub>2</sub>, ( $\Lambda, \blacktriangle$ ), 1000 ppm N<sub>2</sub>O+5% H<sub>2</sub>O; ( $\circ, \bullet$ ), 1000 ppm N<sub>2</sub>O+10% O<sub>2</sub> +5% H<sub>2</sub>O in He. The open symbols refer to CO<sub>3</sub>O<sub>4</sub>, and the closed symbols to the Zn<sub>0.36</sub>Co<sub>0.64</sub>Co<sub>2</sub>O<sub>4</sub> catalysts [17].

#### $N_2O$ decomposition in the presence of excess $O_2$ and $H_2O$

The presence of excess  $O_2$  has a moderate degree of inhibition of the degradation of  $N_2O$ , which indicates that the  $N_2O$  and the  $O_2$  competes for adsorption on the active sites of the catalyst.

The presence of water, however, severely inhibited the reaction of decomposition of  $N_2O$ , possibly due to the hydroxylation of active sites and the adsorption of these. As the temperature increases, water is desorbed from the surface active sites and dehydroxidate thus increasing the efficiency of the N<sub>2</sub>O decomposition. Although the presence of excess  $O_2$  and water inhibited the

reaction, the conversion of  $N_2O$  reaches up to 100% even in the presence of 10 vol %  $O_2$  and 5 vol % water at 300 ° C.

Figure13 shows the effect of the presence of excess  $O_2$  and  $H_2O$  on catalytic activity over  $Zn_{0.36}Co_{0.64}Co_2O_4$  and pure  $Co_3O_4$ .

As the temperature increased, water was desorbed from the surface active sites and the active sites dehydroxylated, the conversion of the N<sub>2</sub>O decomposition increased sharply. The influence of hourly space velocity on the catalytic performance of  $Zn_{0.36}Co_{0.64}Co_2O_4$  at 200°C and the temperature of 100% N<sub>2</sub>O conversion (T100) was also carried out in the Figure 14:



Figure 14: The effect of space velocity on the conversion of  $N_2O$  at 200°C and T100 (the temperature of 100%  $N_2O$  conversion) over  $Zn_{0.36}Co_{0.64}Co_2O_4$  catalyst [17].

The results indicated that the conversion of N<sub>2</sub>O slowly decreased with increasing of space velocity. The N<sub>2</sub>O conversion reached 100% and 71.3% at the space velocity of 15,000 and 55,000  $h^{-1}$  at 200°C, respectively. The temperature of 100% N<sub>2</sub>O conversion was 263 °C at a space velocity of 55,000  $h^{-1}$ . It indicated that the Zn<sub>0.36</sub>Co<sub>0.64</sub>Co<sub>2</sub>O<sub>4</sub> spinel oxide is highly active for N<sub>2</sub>O decomposition reaction and can be operated at a relatively higher space velocity.

#### **Application**

The present results have demonstrated that the introduction of Zn into the spinel structure of

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Co<sub>3</sub>O<sub>4</sub> significantly promoted the catalytic activity for N<sub>2</sub>O decomposition.

The yields are shown in the table 6 at various conditions.

Rh- or Ru-exchanged zeolites and calcined hydrotalcites are the best catalysts for  $N_2O$  decomposition in literature. Compared to the catalysts (Table 6 ),  $Zn_{0.36}Co_{0.64}Co_2O_4$  catalyst is more active for  $N_2O$  decomposition reaction under similar reaction conditions.

**Table 6:** Comparison of catalytic activity of  $Zn_{0.36}Co_{0.64}Co_2O_4$  with the best catalysts in literature<br/>for N<sub>2</sub>O decomposition.

Catalyst_formula <sup>a</sup>	200 °C	250 °C	300 °C	350 °C	400 °C	450 °C
	N <sub>2</sub> O conversion %	N <sub>2</sub> O conversion %	N <sub>2</sub> O conversion %	N <sub>2</sub> O conversion %	N <sub>2</sub> O conversion %	N <sub>2</sub> O conversion %
$Zn_{0.36}Co_{0.64}Co_2O_4^{\ a}$	72(35 <sup>b</sup> )	$98(59^{b},8^{c})$	$100(91^{\rm b},57^{\rm c})$	$100(100^{b},98^{c})$	-	-
Co-Rh-Al-HT <sup>d</sup>	I	-	$100(7^{f})$	-	$100(67^{f})$	$100(100^{\rm e}, 100^{\rm f})$
Rh(Ru)-ZSM-5 <sup>d</sup>	-	20	74	98	-	-

<sup>a</sup> Test conditions:1000ppm N<sub>2</sub>O in He,  $GV = 55,000 \text{ h}^{-1}$ .

<sup>b</sup> With 10% oxygen.

<sup>c</sup>. With 10% oxygen and 5% water.

<sup>d</sup>990 ppm N<sub>2</sub>O in He,  $GV = 60,000 \text{ h}^{-1}$ 

<sup>e</sup> With 2.5% oxygen

<sup>f</sup> With 2.5% oxygen and 2.5% water

The appropriate catalyst is prospective to be used in practice for the removal of N<sub>2</sub>O.

## Best case study 4: Mullite-supported Rh catalyst.

For rhodium catalysts, oxygen desorption is generally considered as the rate limiting step in related reactions of  $NO_x$ . In this case study it was investigated the decomposition of  $N_2O$  over rhodium catalysts [25]-[29]. Many of them suffer from inhibition by oxygen. Different theories exist to explain the fact that rhodium catalysts can decompose  $N_2O$  effectively in oxygen-rich feeds, but are rapidly deactivated because of the saturation of the active centers by adsorbed oxygen.

### <u>Catalyst</u>

From the Rh-team catalysts for the N<sub>2</sub>O decomposition the most widely used, commercially available, is the one supported on  $\gamma$ -Al<sub>2</sub>O<sub>3</sub>.

The activity of the catalyst depends largely on the properties of the carrier, as the  $\gamma$ -Al<sub>2</sub>O<sub>3</sub> contains different amounts of Na ions, different each time modifying the crystalline structure of the surface, thus affecting the active phase of catalyst. During the decomposition of N<sub>2</sub>O mechanism, the desorption of O<sub>2</sub> is the limiting step of this reaction and it has emerged that is a deterrent function of rhodium catalysts, and created strong beams-Rh, blocking the active sites.

X. Zhao et al. in 2009 investigated the activity of a promising material for the  $N_2O$  decomposition due to its capacity to survive in the required high temperature environment or capable to initiate decomposition at low enough temperatures [29]. So a novel catalyst was designed by supporting rhodium on mullite, with the aim of developing a practically capable catalyst for  $N_2O$  used as a propellant, and probing into the role of mullite. Rhodium was chosen as the active component for the reason that Rh has exhibited excellent activities for the abatement of  $N_2O$  at low concentration [29].

#### **Materials and Reaction System**

Mullite is a ceramic material with excellent thermal and mechanical properties: high porosity, thermal stability and unusual crystalline structure. The porous structure can be easily achieved by incorporating appropriate factors or through the decomposition of  $Al(OH)_3$  during the manufacturing process. The oxygen vacancies are abundant due to the balancing of double substitution of Si<sup>+4</sup> and Al<sup>+3</sup>. Mullite synthesized by 2SiO<sub>2</sub>.3Al<sub>2</sub>O<sub>3</sub> using commercial Al (NO<sub>3</sub>)<sub>3</sub> and (TEOS: tetraethylorthosilicate) calcinated at 1473 K (Rh / M). The alumina prepared by  $Al_2O_3$ , calcinated at 773 K (Rh/Al). The Rh content was 5 wt%. For the catalytic preparation wet impregnation and ignition was used at 1473K: [Rh / M-HT] and [Rh / Al-HT].

The catalytic performance of N<sub>2</sub>O decomposition was evaluated in a fixed bed flow reaction system. The reacting gas contained 30 v/v% N<sub>2</sub>O in Ar, and was introduced at a flow rate of 50 mL min<sup>-1</sup>, corresponding to a gas hourly space velocity of 30 000 mL g<sup>-1</sup> h<sup>-1</sup>; and the catalysts were reduced with H<sub>2</sub> at 773 K for 2 h before reaction.

#### <u>Analysis</u>

Figure 15 compares the catalytic activities of different catalysts as a function of reaction temperature. As expected, the Rh/M catalyst exhibited a superior high activity in N<sub>2</sub>O.



Figure 15: N<sub>2</sub>O decomposition on Rh-catalysts [29].

For the Rh / M case: The catalytic activity starts at 473K and at the temperature of 573 K  $(300^{\circ}C)$  achieved the 90% conversion. For the Rh / Al case, the 100% conversion achieved at 773 K while the active surface is 5.7 times greater than that of Rh / M, the rhodium content is the same for both types of catalysts. The good catalytic performance of Rh / M is related to the mechanism of decomposition of N<sub>2</sub>O. In the second step of the mechanism of the decomposition of N<sub>2</sub>O, N<sub>2</sub>O decays via the redox mechanism of oxygen desorption from the catalyst surface. In this catalyst, there are many places covered by oxygen, which may indicate that the adsorption of oxygen is related to the structure of mullite where the loads balanced on a Si<sup>+4</sup> and Al<sup>+3</sup>.

The extra ignition in catalyst at 1473 K shows lower activity, which is associated with the change of crystal structure and the gasification of Rh compounds [29].

In Figure 15 the catalytic activities of different catalysts are compared as a function of reaction temperature. As expected, the Rh/M catalyst exhibited a superior high activity in  $N_2O$  decomposition. The Rh/Al sample also showed a good activity, decomposing  $N_2O$  completely above 773 K. The BET surface area of the Rh/M is 5.7 times lower than that of the Rh/Al, while the Rh contents were approximately equal. But the BET surface area and the active component content are not primary parameters for the Rh/M. The superior activity of the Rh/M can also be confirmed by comparing its activity with that of the Ir-hexaaluminate catalyst as plotted in the Figure. The catalysts calcined at 1473 K demonstrated a significant decrease in the  $N_2O$ 

decomposition reactivity. However, the Rh/M-HT activity was decreased relatively little, revealing its fairly good thermal stability after high-temperature calcination, while the Rh/Al-HT exhibited almost zero activity after calcining at 1473 K. The drastic decrease in N<sub>2</sub>O conversion of the Rh/Al-HT is believed to be correlated with the distinct crystal transition and serious volatilization of Rh species. Moreover, mixed Al<sub>2</sub>O<sub>3</sub>–SiO<sub>2</sub> oxides with the same Al/Si ratio as mullite were also synthesized with the same procedure. The Rh/AS catalyst exhibited a similar activity as that of Rh/Al, while Rh/AS-HT showed only a minimal decease in activity. This can be explained by the formation of the mullite structure.

#### **Application**

The best catalytic behavior is presented by the Rh / M: at the temperature of 573 K (300  $^{\circ}$ C) achieved the 90% conversion of N<sub>2</sub>O.

However, the existing catalysts for the  $N_2O$  decomposition have been found to be either unsurvivable in the required high-temperature environment or unable to initiate decomposition at low enough temperatures. The novel catalyst mullite-supported Rh is highly desirable to be more active and stable to trigger the decomposition of  $N_2O$ .

## <u>Innovative case study: N<sub>2</sub>O-mediated propane oxidative dehydrogenation over</u> <u>Fe-zeolites.- TEOM studies for continuous propylene production in a</u> <u>cyclically-operated reactor.</u>

Steam-activated iron-zeolites (FeZSM-5, Fe-silicalite, Fe-beta) are highly efficient catalysts for the oxidative dehydrogenation of propane using N<sub>2</sub>O, with initial propylene yields up to 25% at 723 K, but they deactivate due to coking. Since catalyst deactivation is reversible, the feasibility of a cyclically-operated process for continuous propylene production, with alternation of reaction and regeneration cycles, has been analyzed. In order to rationally design the sequence in the cyclic experiments, a tapered element oscillating microbalance (TEOM) coupled to on-line analysis of products has been applied. The TEOM provides information on activity, deactivation, and regeneration by simultaneous measurement of reaction and mass changes. In a system with two fixed-bed reactors in parallel using FeZSM-5, stable propene yields higher than 20% were obtained during cycling for 2000 min, using reaction and regeneration temperatures of 723 and 823 K,

respectively. This catalytic process leads to the simultaneous valorization of  $C_3H_8$  and  $N_2O$  and can be economically applied for on-site propylene production using a low-cost source of  $N_2O$ , e.g. in tail-gases of chemical processes where  $N_2O$  is produced in high concentration [30].

## $C_3H_8 + N_2O \rightarrow C_3H_6 + N_2 + H_2O$

Propylene production is very important for further production of polypropylene.

The above reaction is essential not only for the reduction of  $N_2O$  but also for propylene and polypropylene production, both of them used in large amounts in the petrochemical and plastics industry.

By further introduction to a membrane separation of  $N_2$  in the production side more and more propylene is going to be produced.

It is also useful to possibly utilize a membrane reactor so that to combine both separation and reaction in one step for better improvements and economical outcomes [43], [44].

Polyolefins such as polypropylene (PP), polyethylene (PE) and their copolymers, poly(butene-1), polyisobutylene, poly(4-methylpentene-1) are high value commodity polymers. They are produced in high capacity annually and used in manufacturing of plastics, medical/industrial parts and tools, chemicals, and household products. Polypropylene alone has a production capacity of about 8.5 million m.t. annually in North American and 44 million m.t. globally. Polyethylene has a higher global capacity in annual production. Both are predicted to grow in the years to come.

The new process proposed here uses the permeable reactor concept (membrane reactor or permreactor) in order to design a catalytic dehydrogenation membrane reactor (CDMR) of paraffins (alkanes) to olefins (alkenes). The process integrates the CDMR with a polymerization reactor (PR) (gas phase or slurry type) for polyolefin production in a two step reactor process. A third vessel, a separator (or permeator) complements the flow chart to separate, in gas phase, the unreacted olefin from the paraffin at the exit of the polymerization reactor. Olefin (e.g., propylene) is fed back into the PR inlet and the paraffin (e.g., propane) is recycled into the CDMR inlet. Thus, the entire process consists of an integrated reaction-separation-recycling system for the production of polymer grade olefins and subsequently of polyolefin products through a first step

catalytic dehydrogenation process of paraffin-feedstocks.

Production of polymer grade olefins through polymerization reactions usually requires pure monomers as feedstocks to avoid catalyst and solvent contamination and rapid loss of catalyst activity during polymerization. Depending on the type of the downstream utilized polymerization reactor and process (e.g., solution, bulk, suspension, emulsion, gas phase) and the type of the polymer chain propagation reactions (i.e., step-reaction, radical-chain (addition), ionic, coordination) the purity, flowrate and concentration of the monomers (olefins) may vary in the reactor feed. These feed parameters for the monomer are coupled with the polymerization temperature, pressure and reactor volume to make for the production of specific polymers within the desired range of molecular weight, structure and properties (i.e., crystallinity, transparency, viscosity, tensile and impact strength). Specifically, most of the aforementioned commercially available polyolefins are produced with coordination type polymerization and catalysts. Both stirred bed (slurry-multiphase type) reactors and fluid bed (gas phase type) reactors can be used in coordination polymerization employing coordinated complex catalysts (of Ziegler-Natta type) or supported metal oxide catalysts.

Hydrocarbon type solvents are used in the slurry process (e.g., hexane, heptane). Commercially established reactors and reactor modifications of both types operate at conditions which usually range from T=50 to 250 °C and P=1 to 30atm. The exothermic heat of the polymerization reaction is removed by cooling the reactor externally or internally (i.e., by vaporizing a suitable solvent or diluent) or by circulating the unreacted gas through external cooling devices. The above described coordination type polmerization reaction can utilize hydrogen product from the CDMR within the reactor as a chain transfer agent to reduce the polymer molecular weight (MW) and achieve branching which contributes to a decreased crystalline product wherein such an end polymer is required.

Moreover, the olefin monomer can be diluted with a paraffin or cycloparaffin during the polymerization in both types of reactors mentioned above. The proposed processes seek to utilize as part of the diluent or solvent in polymerization reactor the unreacted (non-consumed) paraffin from the CDMR (e.g., propane for polypropylene, ethane for polyethylene, n-butane for poly(butene-1, etc.) to increase the process efficacy and economy. The unreacted paraffins can

remove the exothermic heat of polymerization by vaporization at the reaction conditions. As an example, propane and propylene have low boiling points and under usual polymerization conditions (25-35atm, 55-70°C) they can be transferred from the liquid to the gas phase through boiling.

Mixing with a higher paraffin such as hexane or heptane in the slurry process is also possible in order to increase the solvent efficiency towards the formed polymer and promote mixing in the reactor. By utilizing propane in polypropylene reactor, the propane/propylene separation cost before polymerization can be eliminated.

Furthermore, the use of propane or the corresponding higher carbon paraffins (e.g., butane, pentane) as solvents in polymerization, minimizes the use of higher priced organic solvents and provides an environmentally benign process modification. However, the non-utilized propane or higher carbon paraffins exiting from the PR have to be recycled into the CDMR in order to keep the operation cost low and the continuous production of propylene or higher olefin monomers as shown in Figure 16. Thus, the solvent-like initial paraffin needs to be separated from the unreacted olefin at the top exit of the polymerization reactor.

The proposed separator/permeator (E) as shown in the flow chart of the Figure can operate based on dense or nanoporous solid type polymer reactive membranes, or liquid membranes, containing activated metal ions such as Cu (copper), Ag (silver), Zn (zinc), Cr (chromium), Fe (iron), Ni (nickel), Co (cobalt). These ions have a degree of affinity to form a transporting complex with the permeating olefin (e.g., propylene, ethylene) and to facilitate its transport via the membrane. Moreover, facilitated transport of olefins via metal coated and metal-ion exchange membranes can be also used as separation candidates in separator E.

The described membrane based separation processes compete well energetically and economically with currently applied separation techniques such as the low temperature hydrocarbon distillation.

Furthermore, our long term goal is the development of more stable membranes to perform this separation. Such new membranes can be based on composite metal-polymer, metal-inorganic, or polymer-inorganic materials, stable at higher temperatures and pressures for the separation of

olefins [44].



Figure 16: Flow sheet of integrated olefin/polyolefin production in a membrane reactor [43, 44]



**Figure 17:** Coke content and  $C_3H_6$  yield vs. time during ODHP with N<sub>2</sub>O over steam-activated Fe-zeolites at 723 K. Conditions: 100 mbar  $C_3H_8$  and 100 mbar N<sub>2</sub>O in He, WHSV = 400, 000 ml  $h^{-1} g^{-1}_{cat}$ , and P = 2 bar [30].

- The initial propylene yield was about 19-22% at 723K, with a degree of conversion of about 45% propane and propylene selectivity of about 50%, with complete conversion of N<sub>2</sub>O.
- The interaction of C<sub>3</sub>H<sub>8</sub> with N<sub>2</sub>O through iron zeolites (O<sub>2</sub> absence) does not lead to propylene production, underlining the essential role of oxygen in this dehydrogenation.

The conversion of  $N_2O$  at 673K is 15%, 25% at 723K and 75% at 773K.

The adsorbed atomic oxygen plays an important role in the regeneration of zeolite and in the oxidation of coke.

• In conclusion: the cyclical process of continuous production of propylene by  $N_2O$  ODHP through Fe-zeolites is extremely effective for the oxidative dehydrogenation of propane, but deactivated due to coke formation. The deactivation process is greatly influenced by the type of host zeolite and the reaction temperature.

• The propylene-N<sub>2</sub>O process competes with a number of other processes for propylene production such as the dehydrogenation of propane in a membrane reactor described above.

## 2.1.1 N<sub>2</sub>O Decomposition through membrane

Measurement and analysis of gas permeability in polymeric membranes is an important research tool in a proper design and selection of operating conditions for a gas separation system. Measurement results produce variations in species permeability as a function of feed composition, driving pressure, temperature and membrane type [31]-[36].



Figure 18: A typical schematic presentation of N<sub>2</sub>O decomposition through membranes [34].



Figure 19: The dependence of separation degree from sampling rate for nitrous oxide – oxygen,  $1 - L_{1 \text{ in}}/L_{1 \text{ out}} = 100$ ,  $2 - L_{1 \text{ in}}/L_{1 \text{ out}} = 70$ ,  $3 - L_{1 \text{ in}}/L_{1 \text{ out}} = 50$  [34].

Membranes are promising materials during the  $N_2O$  decomposition and separation, their utilization is not only of interesting because of their ecological aspect but also because of their economical method due to their low cost. Such an example, characterized by low consumption of both materials and energy, is represented below through the Figure 18.

The dependence of separation degree of nitrous oxide from nitrogen and oxygen at different values of feed and flows ratios is presented in Figure 18, respectively.

As it is shown in Figure 19, it is possible to achieve high purity level of nitrous oxide even at small sampling rates. For example, at sampling rate equal to 15%, the concentration of oxygen impurity for the system nitrous oxide – oxygen is decreasing more than 1000 times. As a result, the oxygen concentration is satisfying the purity requirements. And for the system nitrous oxide – nitrogen, the purity requirement is met at a lower sampling rate (not more than 10%).

It was shown that a membrane module with a feeding reservoir might be effectively used for nitrous oxide high purification from nitrogen and oxygen at room temperature and with small product losses [34].

determined as faile of permeabilities.						
Membrane type	α: N <sub>2</sub> O-O <sub>2</sub>	α: N <sub>2</sub> O-N <sub>2</sub>	α: O <sub>2</sub> -N <sub>2</sub>	$\frac{P \times 10^{14}}{\text{mole m/(m}^2 \text{sPa})}$		
Lestosil (present work)	$5.0 \pm 0.2$	$10.1 \pm 0.4$	$2.0 \pm 0.01$	$98.6 \pm 2.8$		
Silicon rubber	109	-	-	204		
Cellulose acetate	6.3	30.4	4.8	592		
Polydimethilsiloxane	7.2	15.5	2.1	145.9		
Perfruoro(2-methylen-4-	0.5	2.2	4.4	7.0		
methyl-1,3-dioxocyclopenante)						
Poly[bis(triflouroethoxy)	6.1	14.6	2.4	72		
Phosphazenel						

 Table 7: Calculated selectivities of nitrous oxide–gas system for various polymer membranes, determined as ratio of permeabilities.

As a membrane module in this case, a commercially available composite non-porous polymeric membrane named "Lestosil" was used, which base is a staircase copolymer of poly(dimethylsiloxane) and poly(diphenylsiloxane). The thickness of the membrane was 5µm.

For permeability measurements a manometric method was used. The values of permeability coefficient were measured for nitrous oxide, nitrogen and oxygen at feed pressure in the range from  $10^5$  to  $4x10^5$  Pa. The value feed pressure was increased and then decreased stepwise. The experimental results and calculated selectivities of the nitrous oxide gas system, determined as ratio of permeabilities, are presented in detail in the Tables 7 and 8.

p <sub>1</sub> x 10 <sup>-5</sup> , Pa	α: N <sub>2</sub> O-O <sub>2</sub>	α: N <sub>2</sub> O-N <sub>2</sub>	$\alpha$ : O <sub>2</sub> -N <sub>2</sub>
1.0	$5.0 \pm 0.2$	$10.0 \pm 0.7$	$2.0 \pm 0.1$
2.0	$4.9\pm0.5$	$9.8 \pm 1.6$	$2.0 \pm 0.3$
3.0	$5.2 \pm 0.6$	$10.5 \pm 1.2$	$2.0 \pm 0.1$
4.0	$5.3 \pm 0.6$	$10.7 \pm 2.1$	$2.0 \pm 0.3$
3.0	$5.2 \pm 0.6$	$10.5 \pm 1.3$	$2.0 \pm 0.2$
2.0	$4.9 \pm 0.4$	$9.8 \pm 1.1$	$2.0 \pm 0.2$
1.0	$4.7 \pm 0.1$	$9.5 \pm 0.7$	$2.0 \pm 0.2$

**Table 8:** Ideal selectivity at different feed pressures.

As it is shown above, the selectivity of systems, consisting of the investigated gases, does not depend on the feed pressure. It is clear that nitrous oxide has the biggest permeability in comparison with its impurity components. In case of one membrane module, the use of the value of enrichment factor is limited to selectivity. Therefore, periodical purification with permeate recycle should be advisably considered to improve the effectiveness of nitrous oxide high purification process. One of the examples of such an operation is the membrane module with a feeding reservoir presented in Figure 18. The purified gas mixture inserts in the feeding reservoir I from where it is actually flown at constant pressure, provided by reducer 2 into the high pressure cavity of membrane module 3. The impurity component is partially removed while the purified mixture which is enriched by the highly permeable component (nitrous oxide) is coming back into feeding reservoir 1 by the vacuum-compressor 4. Then the process is continuously repeated until to reach a high purity level of nitrous oxide. The separation factor F of the process is given by the relationship below:

$$F^{-1} = \frac{C_{1out}}{C_{1in}} = \left(\frac{L_{1in}}{L_{1out}}\right)^{\frac{\alpha^* - 1}{\alpha^*}}$$
(1)

where  $C_{1in}$ : impurity concentration of feed flow of membrane module, ppm;

C<sub>1out</sub>: impurity concentration of retentate flow, ppm;

L<sub>1in</sub>/L<sub>1out</sub>: ratio of feed and retentate flows of membrane module;

 $\alpha^*$ : effective selectivity;

The effective selectivity for the low permeable component is defined as:

$$\alpha^* = \alpha - \frac{p_2}{p_1} (\alpha - 1), \tag{2}$$

where  $p_1$ ,  $p_2$  are pressure of high and low pressure cavities respectively, in Pa;  $\alpha$ : ideal selectivity. The purification degree (f) in this case should be calculated by the following equation:

$$f = \frac{C_0}{C} = \left(\frac{N_0}{N}\right)^{\frac{1-F}{F}}$$
(3)

where C,  $C_0$  and N,  $N_0$  are the quantity of mixture and its concentration, in the feeding reservoir, before and after purification, respectively.

#### 3. Conclusions

The purpose of this review study is to record catalytic processes of decomposition of  $N_2O$ , and propose some innovative solutions that could be investigated and implemented in the future as new technologies. The total contribution of  $N_2O$ , because of its long life time, to the greenhouse effect is 310 times larger than that of the combined  $CO_2$  and  $CH_4$ . The proposals therefore, through this work are as follows:

A number of CDM projects is under development globally for the effective decomposition of  $N_2O$  and the alleviation of its greenhouse effects.

 $Pd/Al_2O_3$  is the best catalyst among other noble metals for the effective decomposition of  $N_2O$ . Oxygen is an inhibitor for the  $N_2O$  decomposition reaction while hydrocarbons and especially propane are catalytic promoters.

Zeolites can play an important role in the  $N_2O$  decomposition. Research in underway to discover efficient zeolite structures to facilitate this decomposition reaction. Fe/ZSM-5 is a good zeolite catalyst to promote the  $N_2O$  decomposition reaction, especially under the presence of NO.

The spinel catalyst  $Zn_{0.36}Co_{0.64}Co_2O_4$  offers complete N<sub>2</sub>O conversion (up to 100%) at temperatures above 250 °C and consists an important class of materials for N<sub>2</sub>O decomposition reaction.

The Rh/Mullite catalyst offers an excellent performance in N<sub>2</sub>O decomposition. This class of catalysts offers 100% conversion of N<sub>2</sub>O at temperatures above 350 °C.

The oxidative dehydrogenation of propane on zeolite catalysts is an effective way to reduce  $N_2O$  into  $N_2$  with the simultaneous production of propylene. The conversion of  $N_2O$  at 400°C reaches 15%, 25% at 450 °C and 75% at 500 °C. However, the process is deactivated due to coke

formation. The deactivation is strongly affected by the type of the host zeolite and the reaction temperature.

The last process competes with the dehydrogenation of propane taking place in a membrane reactor for the production of propylene and hydrogen. Propylene is subsequently used in a polymerization reactor for the production of polypropylene. This process uses membranes and recycling streams to achieve its goals of high yields in propylene and polypropylene production.

Membranes of polymer nature can be used in modules for the effective purification of nitrous oxide from nitrogen and oxygen based on the permeability differences of these three components. It is shown that the selectivity of the permeating gases in the polymer system used (Lestosil) does not depend on the feed pressure.

Use of the catalytic membrane reactor concept where the reaction of propane with  $N_2O$  takes place. This application can be lucrative in the possible separation of propylene product from propane and the other components in order to have an increased consumption of  $N_2O$ .

Thus, the catalytic reaction of propane and nitrous oxide to propylene, provides a double and important application of useful utilization of both  $N_2O$  reduction and propylene production for further use as polypropylene feedstock.

The combination of the above findings can be further investigated and perhaps propose new innovative solutions that will benefit both economically and in terms of energy and materials conservation the decay processes of  $N_2O$ .

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